

# Root nodule bacteria isolated from *Lotus uliginosus* for future use in phytostabilization of arsenic contaminated soils

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## Abstract

In recent decades there has been growing concern around heavy metals and metalloid contamination in soil. Arsenic (As) is a ubiquitous trace metalloid. The high levels of this metalloid in soils are a consequence of human activities and also from natural inputs. In general, the biodiversity of microorganisms and plants decreases drastically in contaminated soils. The knowledge that some leguminous plants, mainly certain species of *Lotus*, are growing well in such soils has attracted our attention for studying symbioses that are well adapted to harsh conditions. In this work we studied the rhizobial population existing in the root nodules of native *Lotus uliginosus* Sch. growing in a central region of Portugal. This legume grows in soils particularly affected by As due the discharge of industrial liquid effluents from fertilizer and chemical facilities. Diversity and tolerance to different concentrations of As of root nodule bacteria were studied. Our results showed that the symbioses between *L. uliginosus* and As tolerant *Bradyrhizobium* isolates were efficient when a nutrient medium containing high As concentrations was used. The present work highlights the capacity of *L. uliginosus* to grow and establish nitrogen-fixing symbioses in soils strongly contaminated with As and its potential for future use to promote vegetation cover to stabilize As contaminated soils.

**Citation:** Soares R, Fareleira P, Colavolpe B, Ruiz OA, Videira e Castro I. 2023. Root nodule bacteria isolated from *Lotus uliginosus* for future use in phytostabilization of arsenic contaminated soils. *Grass Research* 3:8 <https://doi.org/10.48130/GR-2023-0008>

## Introduction

The contamination of ecosystems with heavy metals has become a concern due to their toxicity to biota and also due to them being permanently immobilised in soil components. Moreover, metals influence soil microbes by decreasing their population size, diversity and biochemical activity, thereby altering the structure of soil microbial communities<sup>[1–3]</sup>. Exposure to metals can also result in the establishment of tolerant microbial populations in soil, with important roles in the ecosystem.

Arsenic (As) is a ubiquitous trace metalloid present in almost all environments and widely distributed in soil and water<sup>[4]</sup>. The amounts of As in non-polluted soils are usually lower than 10 mg·kg<sup>-1</sup><sup>[5]</sup>. The occurrence of high levels of As in soils results from human activities, such as mining and smelting processes, in addition to the use of As-based compounds such as insecticides, herbicides, fungicides, algaecides, sheep dips, wood preservatives, dyestuffs and feed additives<sup>[5]</sup>. Natural sources of As comprise volcanic activities, wind-born soil particles, sea salt sprays and microbial volatilization<sup>[6,7]</sup>.

Currently available tools for the remediation of soils polluted with metals or metalloids are costly, time consuming, can be dangerous to people and result in the production of secondary waste<sup>[8,9]</sup>. Phytoremediation has become known as a promising eco-remediation technology by which plants and their associated microbiota are used to eradicate contaminants from polluted soils<sup>[10,11]</sup>, acting through phytoextraction, stabiliza-

tion, immobilization, volatilization and rhizofiltration<sup>[12–15]</sup>.

Soils contaminated by heavy metals frequently have low percentages of organic matter and available nitrogen, limiting plant development<sup>[16]</sup>. Legume plants and their associated root nodule bacteria (rhizobia) are essential components of the biogeochemical cycles in both natural and agricultural ecosystems. We consider that this association could aid rehabilitation of disturbed areas by adding fixed atmospheric nitrogen that can be used by other plants (non-legumes). However, the growth and development of many plant species, including legumes, is affected by the occurrence of high amounts of toxic chemical elements such as As.

Compared to other legumes, *Lotus* species have a higher potential for adaptability to abiotic stresses, frequently surviving in adverse conditions such as those found in polluted soils<sup>[17,18]</sup>. *Lotus* is a cosmopolitan genus with two main centers of diversity, the Mediterranean region (including portions of Europe, Africa, and Western Asia) and Western North America<sup>[19,20]</sup>. The genus *Lotus* comprises about 100 annual and perennial species<sup>[18]</sup>. However, only a small number of *Lotus* species, mainly of agronomic interest, have been studied in relation to their symbionts<sup>[21,22]</sup>. The species of *Lotus* that have been domesticated and improved by selection and plant breeding are: *L. corniculatus*, L.; *L. uliginosus* Schkuhr., also denoted as *L. pedunculatus*; *L. glaber* Mill., also denoted as *L. tenuis*; *L. subbiflorus* Lagasca and *L. ornithopodioides*<sup>[18]</sup>. On the other hand, *L. japonicus* is used as a model for genetic and molecular studies<sup>[23–25]</sup>.

*L. uliginosus* is a perennial legume also used as a tropical forage with a high productive capacity and is recommended for extensive livestock areas in countries of the South American cone and especially in Chile, Brazil and Uruguay. This species was also naturalised in Argentina. Its use is in cattle and sheep rearing and fattening processes and also for soil remediation. It thrives in varied conditions, especially low fertility and moisture as well as soil acidity. It is a species with high spring-summer-autumn forage potential comparable to other traditionally used legumes. This specie has also a outstanding nutritional value, with the presence of condensed tannins, giving it additional nutritional advantages<sup>[26]</sup>. Due to its importance several authors have studied the diversity and phylogenetic relationships of root nodule bacteria of *L. uliginosus* collected from fields in different countries, such as Uruguay<sup>[27]</sup>, Portugal<sup>[28]</sup> and Belgium<sup>[29]</sup>. These authors indicated that those strains were affiliated with *Bradyrhizobium japonicum* or *Bradyrhizobium* sp.

The goal of this research was to investigate a symbiotic legume system effective in nitrogen fixation in a polluted site for possible use during remediation of such soils. In view of the future use of legumes for soil improvement, it is essential to assess the effect of As on the functioning of the symbiosis. Therefore, the objective of this work was also to study the rhizobial population associated with *L. uliginosus* growing in As contaminated soils with regards to genetic diversity, efficiency of nitrogen-fixation and tolerance to As (Supplemental Fig. S1).

## Materials and methods

### Isolation of bacteria from *Lotus uliginosus* root nodules

Bacteria were isolated from root nodules of *Lotus uliginosus* growing under field conditions on an industrially contaminated soil, in the centre of Portugal (GPS: 40.778978329572496, -8.59899224806268), containing high levels of arsenic ( $1.5 \times 10^3$  mg·kg<sup>-1</sup>). Six plants were collected and 5–6 nodules were randomly taken from each plant. The nodules were carefully detached from the roots, surface sterilized (0.25% solution of HgCl<sub>2</sub>)<sup>[28]</sup> and washed extensively with sterile water. Afterwards, nodules were individually crushed and spread on a plate with yeast-mannitol agar (YMA) supplemented with congo red dye<sup>[30]</sup>. Plates were incubated at 28 °C in the dark for 7–10 d. The isolate purity was checked by examining colony morphology and congo red absorption. Isolates were also cultivated in yeast-mannitol agar (YMA) supplemented with bromothymol blue (BTB)<sup>[30]</sup>. Isolates were kept at 4 °C.

### Genetic diversity

ERIC (Enterobacterial Repetitive Intergenic Consensus) - PCR<sup>[31]</sup> was used to assess the genetic diversity of 22 isolates obtained from nodules of *L. uliginosus* as described above. This procedure is used to distinguish strains that are taxonomically very close and has been recognized as appropriate for assessing rhizobial diversity<sup>[32,33]</sup>. DNA was extracted from bacterial liquid cultures using the Aqua pure genomic DNA extraction kit from Bio-Rad, following the kit protocol specifications. ERIC fingerprints were generated using the primers ERIC1R and ERIC2 previously reported by de Bruijn and Versalovic et al.<sup>[31,34]</sup>. Matrices of the Dice coefficient were calculated and cluster analysis was performed using the UPGMA (Unweighted Pair Group with Arithmetic Average) algorithm and the program Free Tree<sup>[35,36]</sup>. The program Tree View (PHILIP) was

used for the construction of dendrograms and evaluation of the respective genetic relationships.

### 16S rRNA region amplification and sequencing

The 16S rRNA region was amplified in seven selected isolates using 41F and 1488R primers as described by Weisburg et al.<sup>[37]</sup>. PCR reactions primers were used at a concentration of 5 μM, together with 15.8 μl of Qiagen kit Taq mix solution (2.5 U of Taq polymerase, 1.5 mM of MgCl<sub>2</sub>, 200 μM of the different dNTP), and approximately 20 ng of genomic DNA, in a final volume of 20 μl. Amplifications were performed following the conditions of Weisburg et al.<sup>[37]</sup>, using an Eppendorf Mastercycler Gradient thermocycler. PCR products of the amplified 16S rRNA region, with an expected size of about 1500 bp, were confirmed by agarose gel electrophoresis. These products were sequenced, with the same primers used for PCR amplification, with an ABI 3730 XL automated sequencer, by STAB VIDA, Caparica, Portugal. Obtained nucleotide sequences were subjected to quality control and edited as necessary using the DNA chromatogram files in Chromas Lite program (version 2.1.1). Homologous sequences were searched in NCBI (National Center for Biotechnology Information) GenBank database ([www.ncbi.nlm.nih.gov](http://www.ncbi.nlm.nih.gov)) using BLASTn tool<sup>[38]</sup>. Nucleotide sequences were aligned using ClustalW within the MEGA 7 platform. A neighbor-joining phylogenetic tree was constructed using the p-distance model and tested with 1000 bootstrap replication within the MEGA 7 platform<sup>[39]</sup>.

### NCBI accession numbers

Sequences obtained previously were deposited in the NCBI database. Below is a list of all accession codes presented next to the correspondent strain:

Isolate 8: OQ145681  
Isolate 10: OQ145686  
Isolate 12: OQ145682  
Isolate 15: OQ145683  
Isolate 21: OQ145684  
Isolate 23: OQ145685  
Isolate 24: OQ145687

### Arsenic tolerance screening of bacterial isolates

Arsenic tolerance of isolates was assessed by evaluating growth inhibition in the presence of different levels of As. For each isolate, a cell suspension was prepared in sterile water and 30 μL of a pre-washed re-suspended aliquot ( $10^6$  cells mL<sup>-1</sup>) was inoculated in 5 mL of yeast-mannitol (YM) liquid medium containing increasing concentrations of arsenic(III) chloride (AsCl<sub>3</sub>) (5, 10, 50 and 200 mg of As mL<sup>-1</sup>). Tubes were kept at 28 °C on an orbital shaker at 100-rev min<sup>-1</sup> for 72 h. Isolate growth was evaluated by a once per day measurement of the optical density at 600 nm. Two replicas for each strain and each concentration were carried out. Finally, the growth of the bacterial isolates was classified into three groups according to the percentage of growth inhibition: sensitive (100%–80%), moderately tolerant (80%–60%) and tolerant (< 60%). Cell suspensions prepared without the addition of As were also included as controls for each isolate.

### Effects of arsenic in the symbiosis

*Lotus uliginosus* cv. Sunrise seeds were surface sterilized and rinsed extensively with sterile distilled water according to<sup>[30]</sup>. Next, seeds were hydrated for 1–2 h in sterilized water and moved to 0.8% w v<sup>-1</sup> agar-water plates for 1–2 days until

germination. The pre-germinated seeds were moved to slants containing 50 mL of N-free Jensen plant medium<sup>[40]</sup>. Seedlings were allowed to establish in this nutrient medium before addition of bacterial inocula and As treatments. Each isolate was inoculated by applying 1 mL of bacterial suspension (approximately  $10^8$  bacterial cells in 1/4 Jensen medium) on the roots of each seedling (3 replicates per isolate). Arsenic was added to liquid Jensen medium (1/4 diluted) as  $\text{AsCl}_3$ , in order to produce concentrations of 0.5, 5, 10, 20, 100 and 200  $\text{mg}\cdot\text{mL}^{-1}$  of As. Plants were supplemented with each As concentration (three replicates per each treatment). Furthermore, controls only with nitrogen (TN) and without nitrogen and not inoculated (T0) were prepared as described by Soares et al.<sup>[22]</sup>. Plants were incubated for six weeks in a controlled-environment chamber with 16 h light/8 h dark cycle at 23 °C (day)/18 °C (night). Plants were observed for nodulation after three and six weeks, and were harvested after six weeks of growth. The colour of the interior of the nodules was also observed and correlated with the presence/absence of leghemoglobin depending on whether they were pink or not, i.e., root nodules coloured pink by leghaemoglobin are caused by a nitrogen-fixing symbiotic relationship between the plant and beneficial bacteria (rhizobia)<sup>[41]</sup>.

Finally, the shoots were dried at 80 °C for two days. The dry weight (aerial biomass) of inoculated plants (X) was used to determine the index of effectiveness (Es) as described by Ferreira & Marques<sup>[42]</sup>. TN and T0 represent the dry weight of plants from nitrogen control and from non-inoculated plants, respectively:  $= \frac{X-T0}{TN-T0} \times 100\%$ .

### Statistical analysis

Data were analysed by one-way analysis of variance (ANOVA) with software STATISTICA version 10 (StatSoft), using the Tukey's honestly significance difference (HSD) test at  $P \leq 0.05$ .

## Results

### Bacterial diversity

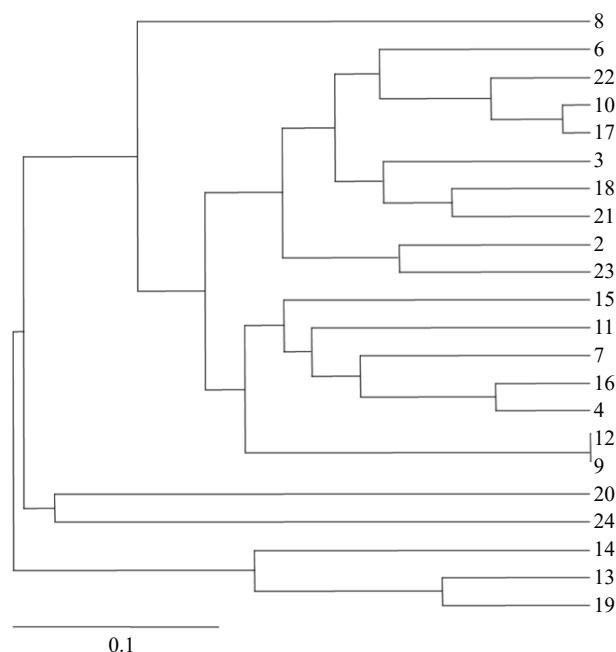
In order to address the objective of this research, we isolated bacterial strains from *L. uliginosus* growing under field conditions in an arsenic contaminated site in the central region of Portugal. In this context, 22 isolates were obtained from root nodules of *L. uliginosus*. All isolates were slow growers, showed little congo red absorption, and had an alkaline reaction on BTB, indicated by a blue colour, which is usually produced by *Bradyrhizobium* spp.<sup>[30]</sup> (Supplemental Fig. S2). Congo red is often incorporated in culture media for isolating rhizobia or for testing the purity of rhizobia cultures. Rhizobia typically do not absorb congo red or absorb it weakly, while other bacteria absorb it strongly.

The assessment of genotypic diversity of the natural population nodulating *L. uliginosus* was achieved by ERIC-PCR. The analysis of the fingerprinting patterns of each isolate showed the existence of several clusters (Fig. 1). This dendrogram was used to determine the similarities among isolates. Results showed the presence of a high genetic diversity in the population, despite the high contamination by As in the original soil. These isolates showed multiple fingerprinting patterns and no single dominant genotype was apparent from our results.

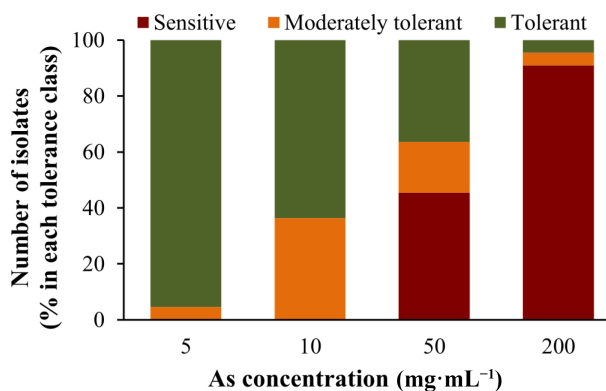
### Arsenic tolerance

Results obtained for As tolerance after 72 h (Fig. 2, Supplemental Fig. S3 and Table 1) showed that, for the highest

concentration tested (200  $\text{mg}$  of  $\text{As mL}^{-1}$ ), most isolates were considered As sensitive (percentage of growth inhibition (GI) of 80%–100%). Only two isolates, 13 and 12, were considered as moderately tolerant (GI 60%–80%) and tolerant (GI < 60%), respectively. On the other hand, for 50  $\text{mg}\cdot\text{mL}^{-1}$  of As about 36% of isolates (isolates 8, 12, 13, 14, 15, 16, 17 and 23) were tolerant, and 18% were considered moderately tolerant (isolates 7, 18, 19 and 22). However, an inverse situation was verified for 10 and 5  $\text{mg}\cdot\text{mL}^{-1}$  of As, where no isolates were found sensitive. While for 10  $\text{mg}\cdot\text{mL}^{-1}$  64% of the isolates were tolerant, for the lowest concentration of As used, 5  $\text{mg}\cdot\text{mL}^{-1}$ , a large majority of isolates (95%) was considered as tolerant.



**Fig. 1** Dendrogram showing the diversity of root nodule bacteria (*Bradyrhizobium* sp.) isolated from *L. uliginosus*, derived from ERIC-PCR fingerprints using UPGMA method, at 85% similarity.



**Fig. 2** Tolerance of the *L. uliginosus* isolates to different As concentrations (5, 10, 50 and 200  $\text{mg}\cdot\text{mL}^{-1}$ ). Isolates were classified according to the percentage of growth inhibition relative to controls grown in the absence of As, being considered sensitive (80%–100%), moderately tolerant (60%–80%) and tolerant (< 60%). Stacked-columns indicate the percentage of isolates in the three tolerance classes for each As concentration.

**Table 1.** Nodulation phenotype of *L. uliginosus* plants inoculated with the different isolates upon different As concentrations.

Isolates	As concentration (mg·mL <sup>-1</sup> )									
	0		0.5		5		10		20	
	3 w	6 w	3 w	6 w	3 w	6 w	3 w	6 w	3 w	6 w
8	+	+	+	+	+	+	+	+	-	-
	+	+	+	+	+	+	+	+	-	-
10	+	+	+	+	+	+	+	+	-	-
	+	+	-	±	-	±	-	-	-	-
12	+	+	+	+	+	+	±	±	-	-
	+	+	+	+	+	+	±	±	-	-
15	+	+	-	+	-	+	+	+	-	+
	+	+	+	+	+	+	+	+	-	+
21	+	+	-	+	+	+	-	+	-	+
	+	+	+	+	+	+	-	+	-	-
23	+	+	-	+	-	+	-	+	-	+
	+	+	-	+	-	+	+	+	-	+
24	+	+	+	+	-	+	-	+	-	-
	+	+	+	+	-	+	-	+	-	-

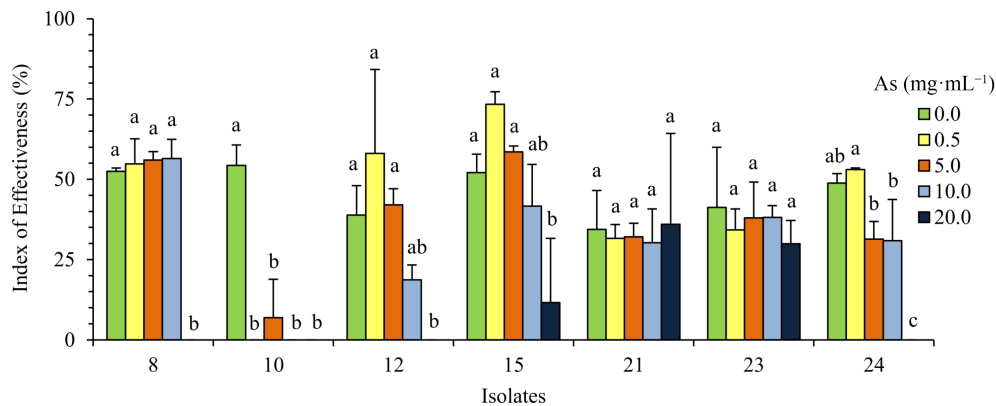
w, incubation weeks; +, presence of nodules; -, absence of nodules; ±, presence of small and white nodules.

**Effects of arsenic on the symbiosis**

For the evaluation of the effects of As on the symbiosis with *L. uliginosus*, seven isolates (namely 8, 10, 12, 15, 21, 23 and 24) belonging to different clusters (Fig. 1) and with different levels of tolerance to As (Supplemental Fig. S3) were chosen to inoculate *L. uliginosus* cv. Sunrise seedlings. Assays were performed with six As concentrations (0.5, 5, 10, 20, 100 and 200 mg of As mL<sup>-1</sup>) and results were recorded three and six weeks after the addition of bacterial inocula and As to the seedlings. Briefly, plants did not tolerate the highest As concentrations used, 100 and 200 mg·mL<sup>-1</sup>, and died one week after As addition. Arsenic toxicity affected the symbiosis and the different isolates also showed different symbiotic performances as the arsenic concentration increased up to 20 mg·mL<sup>-1</sup> (Fig. 3).

One of the effects of As in the symbiosis was demonstrated by the delay in nodulation (Table 1), especially for higher concentrations of As tested, when compared to the controls without the addition of As. A more drastic effect was the impairment of the symbiosis by the lack of nodulation. For the highest As concentration allowing plant survival, 20 mg·mL<sup>-1</sup>, only plants inoculated with isolates 15, 21 and 23 presented pink root nodules (i.e., functional) at least in one of the replicates. In this case, nodulation was delayed, but the symbiosis remained effective in some replicates, depending on the isolate tested. Plants inoculated with isolates 8 and 12 did not show delayed nodulation but nodules were only formed at As concentrations up to 10 mg·mL<sup>-1</sup>. However, at this concentration plants inoculated with isolate 12 had small and white nodules (i.e., ineffective). Nodulation was also observed (with the presence of pink nodules) in plants inoculated with isolate 24 until 10 mg of As mL<sup>-1</sup> but with a delay of 3 weeks. Plants inoculated with isolate 10 and growing with 0.5 and 5 mg of As mL<sup>-1</sup> showed also a delay in nodulation and nodules were small and white. This nodulation phenotype indicates an ineffective symbiosis, inversely to what happened with plants grown without the addition of As which had pink nodules.

These plant-inoculation experiments using several As concentrations were also performed to evaluate the symbiotic effectiveness, by determining the shoot dry weight of *L. uliginosus* plants after 6 weeks of growth. The As concentrations used in these experiments were higher than those usually present in groundwater used for irrigation and in soils in various countries, e.g. 3,10 µg·L<sup>-1</sup> and 22 mg·kg<sup>-1</sup>, respectively, in Argentina, or 3,700 µg·L<sup>-1</sup> and 196 mg·kg<sup>-1</sup>, respectively, in India<sup>[43]</sup>. All the isolates chosen for these experiments were considered effective in the absence of As and the respective indices of effectiveness (Es), under these conditions, ranged between 35% and 55% (Fig. 3). These isolates also showed different levels of symbiotic performance and indices of effectiveness upon different As concentrations. In general, results were congruent with the observed nodulation phenotypes (Table 1). Isolates 15, 21 and 23 were able to establish an efficient symbiosis with the host plant at 20 mg of As mL<sup>-1</sup>. Plants inoculated with isolate 15 showed an increase in the aerial part and the consequent increase in the indices of effectiveness at the concentration of 0.5 mg As mL<sup>-1</sup>, which was near 75% and the highest Es of these experiments. This



**Fig. 3** Effect of As on the symbiosis. Values represent the average of the index of effectiveness (Es) ± SD of *L. uliginosus* plants inoculated with the bacterial isolates and grown with different As concentrations (0, 0.5, 5, 10 and 20 mg·mL<sup>-1</sup>). Different letters express significant differences between plants inoculated with each isolate at several As concentrations according to Tukey's HSD test at  $P < 0.05$ .

strain was therefore considered highly effective in the presence of As ( $0.5 \text{ mg}\cdot\text{mL}^{-1}$ ). However, for the other two isolates, 21 and 23 respectively, no significant differences were found between the various treatments, including the control without As addition, and their respective indices of effectiveness were lower but still considered as effective in nitrogen fixation. On the other hand, isolates 8, and 24, were able to establish an efficient symbiosis until the concentration of  $10 \text{ mg}$  of As  $\text{mL}^{-1}$ . For the first isolate, no significant differences were found between the various treatments ( $0.5$ ,  $5$  and  $10 \text{ mg}$  As  $\text{mL}^{-1}$ ) including the control without addition of As. The indices of effectiveness of plants inoculated with isolate 12 and with  $0.5$  and  $5 \text{ mg}$  of As  $\text{mL}^{-1}$  did not show significant differences between each other and the control (without As addition). However, for  $10 \text{ mg}\cdot\text{mL}^{-1}$  of As the index of effectiveness was very low (19%) and the symbiosis was considered as ineffective, the plants presenting small and white nodules as previously mentioned. Lastly, plants inoculated with isolate 10 had a very weak performance in the presence of As (with plants having small and white nodules) and were only able to establish an efficient symbiosis when As was completely absent, showing indices of effectiveness significantly different from all the remaining treatments.

Interestingly, among the isolates that were able to establish an efficient symbiosis while sustaining the highest As concentration tolerated by plants,  $20 \text{ mg}\cdot\text{mL}^{-1}$  (isolates 15, 21 and 23), isolate 15 was also among those with higher indices of effectiveness in the absence of As. Moreover, this isolate had the highest Es shown, near 75% at  $0.5 \text{ mg}\cdot\text{mL}^{-1}$  of As being considered as highly effective in nitrogen fixation.

### 16S rRNA region phylogeny

Aligned sequences of the partial 16S rRNA region were used to construct the phylogenetic tree shown in Fig. 4. The isolates from this study were all clustered with *Bradyrhizobium* spp. Sequences obtained from isolates 8, 10, 15, 23 and 24 shared 100% sequence identity and the closest strains were *Bradyrhizobium* spp. isolated from *Lotus uliginosus* (*Bradyrhizobium* sp. 3LBC, 8LBI, SEMIA 839 and NZP2309<sup>[21,28]</sup>), *Cytisus triflorus* (*Bradyrhizobium* sp. CTS12), *Cytisus scoparius* (*Bradyrhizobium* genosp. AD Cs6020<sup>[44]</sup>) *Vigna unguiculata* L. (*Bradyrhizobium* sp. VUPMI37) and *Ulex europaeus* (*Bradyrhizobium* sp. ICMP 12674<sup>[45]</sup>). Isolates 12 and 21 shared 100% sequence identity and the closest strains were *Bradyrhizobium* spp. isolated from *Glycine max* (*Bradyrhizobium japonicum* GI-4 and J5 and *Bradyrhizobium* sp. 323S2<sup>[46,47]</sup>), *Pigeonpea* (*Bradyrhizobium* sp. RP6) and *Erythrina brucei* (*Bradyrhizobium shewense* ERR11T<sup>[48]</sup>).

### Discussion

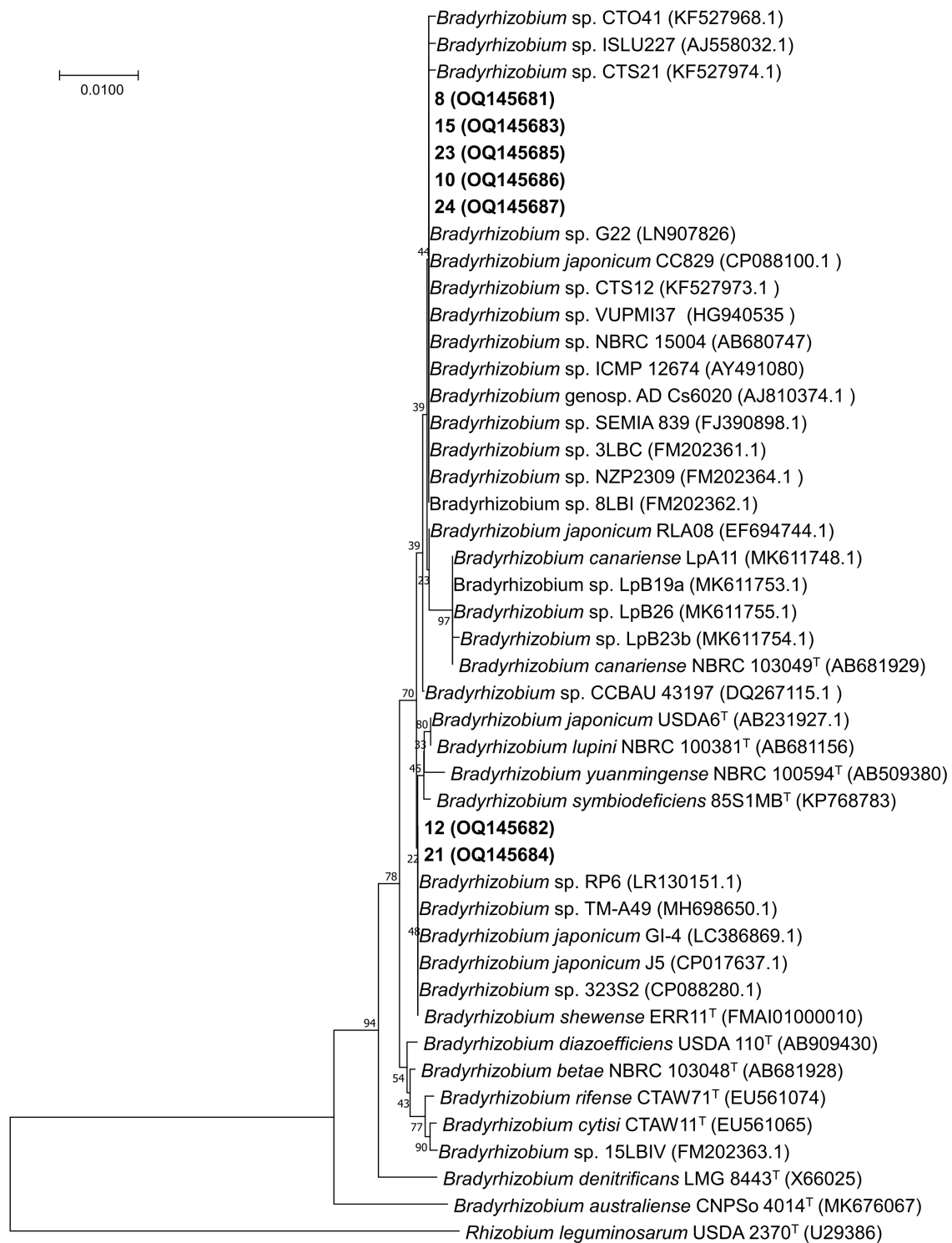
Biological nitrogen fixation, including the contribution made by symbioses in the root nodules of legumes, supplies a large proportion of the nitrogen that increases soil fertility in natural and agro-ecosystems<sup>[49]</sup>. In recent years, the selection of symbiotic or free-living plant growth promoting rhizobacteria with remediation abilities has gained prominence, since such strains could help plants to grow in polluted soils or could even limit the inclusion of contaminants into plant tissues<sup>[50]</sup>. In particular, the rhizobium-legume symbiotic interaction has been highlighted as a promising tool to be used for bioremediation of As and heavy metals in soils<sup>[51–55]</sup>.

It has been reported by several authors that associations between plants and microorganisms increase the bioremediation potential of plants<sup>[56–58]</sup>. In this study we found a large diversity of root nodule bacteria isolated from *L. uliginosus* growing in soils contaminated with As in Portugal. These results are different from previous data indicating a lack of genetic diversity in the population of *Rhizobium* isolated from *Trifolium* sp. in an analogous area contaminated by heavy metals<sup>[59,60]</sup>. However, similar results have been reported by Carrasco et al.<sup>[51]</sup> and also by Delorme et al.<sup>[61]</sup>, who observed that the presence of rhizobial genotypes in soils with high heavy metal concentration were not different from those existing in soils with low metal levels. These authors showed that the rhizobial population was very diverse and the presence of heavy metals did not lead to a decrease in diversity. According to Rangel et al.<sup>[62]</sup>, a wide diversity of bacteria resistant to As was verified, including several rhizobial genera such as *Azorhizobium*, *Mesorhizobium*, *Rhizobium*, *Burkholderia* (now either in *Paraburkholderia* or *Trinickia*)<sup>[63]</sup>. Therefore, biological nitrogen fixation has a great potential to be used in the future for phytostabilization purposes, given the high number of host legume species of the mentioned rhizobia that can be tested in the field and also because legumes accumulate heavy metals mainly in roots and show a low level of metal translocation to the shoot<sup>[55]</sup>.

In this study, we have isolated bacteria from root nodules of *L. uliginosus* growing in soil contaminated by arsenic. These bacteria were morphologically characterized, by growth rate and reaction on BTB, as identical to those strains belonging to the genus *Bradyrhizobium*. Moreover, *L. uliginosus* root nodule bacteria seem to be constrained to *B. japonicum* and *Bradyrhizobium* sp. symbionts, rarely harbouring endophytic bacteria<sup>[27,28]</sup>. These results were confirmed for a set of seven isolates (namely, 8, 10, 12, 15, 21, 23 and 24) used to test the effects of As in the symbiosis, which were molecularly identified as *Bradyrhizobium* spp., using 16S rRNA region sequencing. A detailed phylogenetic analysis comprising appropriate molecular markers<sup>[22]</sup> could reveal if these isolates represent novel species or described ones, and compare them with the ones already described taxonomically from *L. uliginosus* that did not originate from As-contaminated soils<sup>[18,24]</sup>.

In the tests performed to assess As tolerance, it was found that all the isolates were able to grow at concentrations of  $5$  and  $10 \text{ mg}$  As  $\text{mL}^{-1}$  revealing higher tolerance than observed by Deepika et al.<sup>[64]</sup> for *Rhizobium radiobacter* (strain VBCK1062) when using  $10 \text{ mM}$  As in YM liquid medium.

In the tests carried out with *L. uliginosus* plants, the effects of As on nodulation and growth were recorded three and six weeks after seeding. The number of nodules and growth of the plants were affected by the presence of As in the medium, producing a severe effect in the early stage of nodulation. This could be due to the effect of As on rhizobia survival and initial tolerance<sup>[65]</sup>. The negative effects of As in the nodulation process found in our study are similar to those described by several authors. In fact, Reichman<sup>[66]</sup> observed that treatment with As induced a significant decrease in the total and effective number of nodules in soybean. Also, Pajuelo et al.<sup>[67]</sup> found important proof of damage by As on roots as well as a reduction of root hair number, which was associated with a decrease of around 90% in rhizobial infection number in plants



**Fig. 4** Phylogeny of the partial 16S rRNA gene with a total of 1189 aligned positions. Confidence bootstrap values are presented near each node. NCBI GenBank accession codes are presented next to each strain. Isolates obtained in this work are in bold. *Rhizobium leguminosarum* USDA 2370T (U29386) was selected as an outgroup.

of *Medicago sativa*. Moreover, the reduction in nodulation observed in legumes exposed to other metals, such as Cu<sup>2+</sup>, Cd<sup>2+</sup> and Hg<sup>2+</sup>, was also attributed to the damage of metals on root hairs<sup>[68]</sup>. On the other hand, Bustingorri et al.<sup>[69]</sup> demonstrated that As causes membrane damage and decreases the chlorophyll content. This metalloid has no known biological

function and fundamentally disrupts the cell growth and metabolism of living cells<sup>[64]</sup>. The adaptive tolerance response occurring in some rhizobial strains to the perturbation caused by As contamination, was also reported by others<sup>[47]</sup>. Our results confirmed that some *Bradyrhizobium* isolates tested for the evaluation of the effects of As in the symbiosis showed

capacity to establish an efficient symbiosis, even in the presence of the highest As concentration (20 mg·mL<sup>-1</sup>), mainly isolate 15. This isolate had high effectiveness when compared to the As-free inoculated controls and to the TN and T0 controls.

Biological nitrogen fixation, including the contribution made by symbioses with legumes, supplies a large proportion of the nitrogen that increases soil fertility in natural and agro-ecosystems<sup>[49]</sup>. In recent years, the selection of symbiotic or free-living plant growth promoting rhizobacteria with remediation abilities has gained prominence, since such strains could help plants to grow in polluted soils or could even limit the inclusion of contaminants into plant tissues<sup>[50]</sup>. In particular, the *Rhizobium*-legume symbiotic interaction has been highlighted as a promising tool to be used for bioremediation of As and heavy metals in soils<sup>[51–54]</sup>. In this study, we have isolated bacteria from *L. uliginosus* with the capacity to establish an efficient symbiosis, even in the presence of high concentrations of As.

We consider *Bradyrhizobium* isolate 15, as the most promising to be tested *in situ* for their applicability for phytoremediation in sites polluted by this metalloid.

## Conclusions

The main impact of this work is the possible use of autochthonous legume plants and their micro-symbionts, such as the symbioses with *L. uliginosus* and *Bradyrhizobium*, for the phytostabilization of contaminated soils, helping its fertilization. Results reveal that root nodule bacteria isolated from *L. uliginosus* growing in polluted soils, mainly those tolerant to the highest concentration of arsenic, can be symbiotically effective upon high As concentrations. This can be a very interesting aspect and can be used in the future, in bioremediation experiments on contaminated soils using native legumes, since it could have a positive ecological impact in those sites. We consider that the dual function of As bioremediation plus soil nitrogen enhancement can be achieved by effective symbiotic nodulation in affected As-metal-soils. Such a tolerant and functional symbiosis can support vegetation cover, stabilizing As-contaminated soils, and consequently it could be a smart practice for phytostabilization.

## Acknowledgments

This work was supported by the cooperation project between Portugal and Argentina: 'The genus *Lotus* and their utilization for the restoration of soils contaminated with heavy metals. The biochemistry and their symbionts', FCT/DREBM 00264, Proc. 4.1.3 and also by the European project: 'Raising the bio-based industrial feedstock capacity of marginal Lands (Margin Up)', n°101082089.

## Conflict of interest

The authors declare that they have no conflict of interest.

**Supplementary Information** accompanies this paper at (<https://www.maxapress.com/article/doi/10.48130/GR-2023-0008>)

## Dates

Received 12 January 2023; Accepted 20 April 2023; Published online 10 May 2023

## References

- Castro-Larragoitia J, Kramar U, Puchelt H. 1997. 200 years of mining activities at La Paz/San Luis Potosi/Mexico — Consequences for environment and geochemical exploration. *Journal of Geochemical Exploration* 58:81–91
- Oliveira A, Pampulha ME, Neto MM, Almeida AC. 2009. Enumeration and characterization of arsenic-tolerant diazotrophic bacteria in a long-term heavy-metal-contaminated soil. *Water, Air, and Soil Pollution* 200:237–43
- Tipayno SC, Truu J, Samaddar S, Truu M, Preem JK, et al. 2018. The bacterial community structure and functional profile in the heavy metal contaminated paddy soils, surrounding a nonferrous smelter in South Korea. *Ecology and Evolution* 8:6157–68
- Chakraborti D, Rahman MM, Chatterjee A, Das D, Das B, et al. 2016. Fate of over 480 million inhabitants living in arsenic and fluoride endemic Indian districts: Magnitude, health, socio-economic effects and mitigation approaches. *Journal of Trace Elements in Medicine and Biology* 38:33–45
- Adriano DC. 2001. *Trace elements in terrestrial environments: biogeochemistry, bioavailability, and risks of metals*. New York: Springer. xii, 867 pp. <https://link.springer.com/book/10.1007/978-0-387-21510-5>
- Nriagu JO. 1990. Global metal pollution: poisoning the biosphere? *Environment: Science and Policy for Sustainable Development* 32:7–33
- Frankenberger WT, Arshad M. 2002. Volatilization of arsenic in environmental chemistry of arsenic. In *Environmental Chemistry of Arsenic*, ed. Frankenberger WT. New York: Marcel Dekker. pp. 363–380.
- Sharma S, Tiwari S, Hasan A, Saxena V, Pandey LM. 2018. Recent advances in conventional and contemporary methods for remediation of heavy metal-contaminated soils. *3 Biotech* 8:216
- Haq S, Bhatti AA, Dar ZA, Bhat SA. 2020. Phytoremediation of heavy metals: an eco-friendly and sustainable approach. In *Bioremediation and Biotechnology*, eds. Hakeem K Bhat R, Qadri H. Cham: Springer. [https://doi.org/10.1007/978-3-030-35691-0\\_10](https://doi.org/10.1007/978-3-030-35691-0_10)
- Ma Y, Oliveira RS, Freitas H, Zhang C. 2016. Biochemical and molecular mechanisms of plant-microbe-metal interactions: relevance for phytoremediation. *Frontiers in Plant Science* 7:918
- Pinto AP, de Varennes A, Dias CMB, Lopes ME. 2018. Microbial-assisted phytoremediation: a convenient use of plant and microbes to clean up soils. In *Phytoremediation: Management of Environmental Contaminants*, eds. Ansari AA, Gill SS, Gill R, Lanza GR, Newman L. Switzerland: Springer International Publishing. Volume 6, pp. 21–87. [https://doi.org/10.1007/978-3-319-99651-6\\_2](https://doi.org/10.1007/978-3-319-99651-6_2)
- Salt DE, Smith RD, Raskin I. 1998. Phytoremediation. *Annual Review of Plant Physiology and Plant Molecular Biology* 49:643–68
- Wenzel WW, Lombi E, Adriano DC. 1999. Biogeochemical processes in the rhizosphere: role in phytoremediation of metal-polluted soils. In *Heavy Metal Stress in Plants: From Molecules to Ecosystems*, eds. Prasad MNV, Hagemeyer J. Berlin, Heidelberg: Springer. pp. 273–303. [https://doi.org/10.1007/978-3-662-07745-0\\_13](https://doi.org/10.1007/978-3-662-07745-0_13)
- Wenzel WW, Adriano DC, Salt D, Smith R. 1999. Phytoremediation: a plant—microbe-based remediation system. In *Bioremediation of Contaminated Soils*, eds. Adriano DC, Bollag JM, Frankenberger WT, Sims RC. New York: Academic Press. pp. 457–508. <https://doi.org/10.2134/agronmonogr37.c18>.

15. Raklami A, Meddich A, Oufdou K, Baslam M. 2022. Plants-microorganisms-based bioremediation for heavy metal cleanup: recent developments, phytoremediation techniques, regulation mechanisms, and molecular responses. *International Journal of Molecular Sciences* 23:5031
16. Hu X, Wang J, Lv Y, Liu X, Zhong J, et al. 2021. Effects of heavy metals/metalloids and soil properties on microbial communities in farmland in the vicinity of a metals smelter. *Frontiers in Microbiology* 12:707786
17. de los Santos GG, Steiner JJ, Beuselinck PR. 2001. Adaptive ecology of *Lotus corniculatus* L. genotypes: II. crossing ability. *Crop Science* 41:564–70
18. Escaray FJ, Menendez AB, Gárriz A, Pieckenstain FL, Estrella MJ, et al. 2012. Ecological and agronomic importance of the plant genus *Lotus*. Its application in grassland sustainability and the amelioration of constrained and contaminated soils. *Plant Science* 182:121–33
19. Howieson JG, O'Hara GW, Carr SJ. 2000. Changing roles for legumes in Mediterranean agriculture: developments from an Australian perspective. *Field Crops Research* 65:107–22
20. Sokoloff DD, Lock JM. 2005. Legumes of the world. tribe loteae. In *Legumes of the World*, eds. Lewis G, Schrire B, Mackinder B, Lock M. Richmond VA: Royal Botanic Gardens. pp. 455–65.
21. Lorite MJ, Estrella MJ, Escaray FJ, Sannazzaro A, Videira e Castro IM, et al. 2018. The rhizobia-*Lotus* symbioses: deeply specific and widely diverse. *Frontiers in Microbiology* 9:2055
22. Soares R, Trejo J, Lorite MJ, Figueira E, Sanjuán J, et al. 2020. Diversity, phylogeny and plant growth promotion traits of nodule associated bacteria isolated from *Lotus parviflorus*. *Microorganisms* 8:499
23. Diaz P, Borsani O, Monza J. 1995. Effect of inoculation and nitrate on nitrate reductase activity and acetylene reduction activity in *Lotus sp.-Rhizobium loti* symbiosis. *Symbiosis* 19:53–63
24. Pajuelo E, Stougaard J. 2005. *Lotus japonicus's* a model system. In *Lotus japonicus Handbook*, ed. Márquez AJ. Netherlands: Springer. pp. 3–24. [https://doi.org/10.1007/1-4020-3735-X\\_1](https://doi.org/10.1007/1-4020-3735-X_1).
25. Mun T, Bachmann A, Gupta V, Stougaard J, Andersen SU. 2016. *Lotus* Base: an integrated information portal for the model legume *Lotus japonicus*. *Scientific Reports* 6:39447
26. Díaz P, Borsani O, Monza J. 2005. *Lotus*-related species and their agronomic importance. In *Lotus japonicus Handbook*, ed. Márquez AJ. Netherlands: Springer. pp. 25–37. [https://doi.org/10.1007/1-4020-3735-X\\_2](https://doi.org/10.1007/1-4020-3735-X_2).
27. Batista L, Tomasco I, Lorite MJ, Sanjuán J, Monza J. 2013. Diversity and phylogeny of rhizobial strains isolated from *Lotus uliginosus* grown in Uruguayan soils. *Applied Soil Ecology* 66:19–28
28. Lorite MJ, Videira e Castro I, Muñoz S, Sanjuán J. 2012. Phylogenetic relationship of *Lotus uliginosus* symbionts with bradyrhizobia nodulating genistoid legumes. *FEMS Microbiology Ecology* 79:454–64
29. De Meyer SE, Van Hoorde K, Vekeman B, Braeckman T, Willems A. 2011. Genetic diversity of rhizobia associated with indigenous legumes in different regions of Flanders (Belgium). *Soil Biology and Biochemistry* 43:2384–96
30. Somasegaran P, Hoben HJ. 1994. *Handbook for rhizobia: methods in legume-rhizobium technology*. New York: Springer. 450 pp. <https://doi.org/10.1007/978-1-4613-8375-8>
31. de Bruijn FJ. 1992. Use of repetitive (repetitive extragenic palindromic and enterobacterial repetitive intergeneric consensus) sequences and the polymerase chain reaction to fingerprint the genomes of *Rhizobium meliloti* isolates and other soil bacteria. *Applied and Environmental Microbiology* 58:2180–87
32. Zhou S, Li Q, Jiang H, Lindström K, Zhang X. 2013. *Mesorhizobium sangaii* sp. nov., isolated from the root nodules of *Astragalus luteolus* and *Astragalus ernestii*. *International Journal of Systematic and Evolutionary Microbiology* 63:2794–99
33. Irisarri P, Cardozo G, Tartaglia C, Reyno R, Gutiérrez P, et al. 2019. Selection of competitive and efficient rhizobia strains for white clover. *Frontiers in Microbiology* 10:768
34. Versalovic J, Koeuth T, Lupski JR. 1991. Distribution of repetitive DNA sequences in eubacteria and application to fingerprinting of bacterial genomes. *Nucleic Acids Research* 19:6823–31
35. Pavlíček A, Hrdá Š, Flegr J. 1999. Free-Tree-freeware program for construction of phylogenetic trees on the basis of distance data and bootstrap/jackknife analysis of the tree robustness. Application in the RAPD analysis of genus *Frenkelia*. *Folia Biologica (Praha)* 45:97–99
36. Efron B. 2003. Second thoughts on the bootstrap. *Statistical Science* 18:135–40
37. Weisburg WG, Barns SM, Pelletier DA, Lane DJ. 1991. 16S ribosomal DNA amplification for phylogenetic study. *Journal of Bacteriology* 173:697–703
38. Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. 2015. Basic local alignment search tool. *Journal of Molecular Biology* 215:405–10
39. Kumar S, Stecher G, Tamura K. 2016. MEGA7:molecular evolutionary genetics analysis version 7.0 for bigger datasets. *Molecular Biology and Evolution* 33:1870–74
40. Jensen HL. 1942. Nitrogen fixation in leguminous plants. I. General characters of root-nodule bacteria isolated from species of *Medicago* and *Trifolium* in Australia. *Proceedings of the Linnean Society of New South Wales* 67:98–108
41. Mandal BK, Suzuki KT. 2002. Arsenic round the world: a review. *Talanta* 58:201–35
42. Ferreira EM, Marques JF. 1992. Selection of Portuguese *Rhizobium leguminosarum* bv. *trifolii* strains for production of legume inoculants. *Plant and Soil* 147:151–58
43. Chatterjee A, Das D, Mandal BK, Chowdhury TR, Samanta G, et al. 1995. Arsenic in ground water in six districts of West Bengal, India: the biggest arsenic calamity in the world. Part I. Arsenic species in drinking water and urine of the affected people. *Analyst* 120:643–50
44. Lafay B, Burdon JJ. 2006. Molecular diversity of rhizobia nodulating the invasive legume *Cytisus scoparius* in Australia. *Journal of Applied Microbiology* 100:1228–38
45. Weir BS, Turner SJ, Silvester WB, Park DC, Young JM. 2004. Unexpectedly diverse *Mesorhizobium* strains and *Rhizobium leguminosarum* nodulate native legume genera of New Zealand, while introduced legume weeds are nodulated by *Bradyrhizobium* species. *Applied and Environmental Microbiology* 70:5980–87
46. Bromfield ESP, Cloutier S, Tambong JT, Tran Thi TV. 2017. Soybeans inoculated with root zone soils of Canadian native legumes harbour diverse and novel *Bradyrhizobium* spp. that possess agricultural potential. *Systematic and Applied Microbiology* 40:440–47
47. Mason MLT, Tabing BLC, Yamamoto A, Saeki Y. 2018. Influence of flooding and soil properties on the genetic diversity and distribution of indigenous soybean-nodulating bradyrhizobia in the Philippines. *Heliyon* 4:e00921
48. Aserse AA, Woyke T, Kyrpides NC, Whitman WB, Lindström K. 2017. Draft genome sequences of *Bradyrhizobium shewense* sp. nov. ERR11<sup>T</sup> and *Bradyrhizobium yuanmingense* CCBau 10071<sup>T</sup>. *Stand Genomic Science* 12:74
49. Crespo D. 2006. The role of pasture improvement on the rehabilitation of the montado/dehesa system and in developing its traditional products. In *Animal Products from the Mediterranean Area*, EAAP publication No. 119, eds. Ramalho Ribeiro J, Horta A, Mosconi C, Rosati A. The Netherlands: Wageningen Academic Publishers, pp. 185–195. <https://doi.org/10.3920/978-90-8686-568-0>.
50. Armendariz AL, Talano MA, Wevar Oller AL, Medina MI, Agostini E. 2015. Effect of arsenic on tolerance mechanisms of two plant growth-promoting bacteria used as biological inoculants. *Journal of Environmental Sciences* 33:203–10



51. Carrasco JA, Armario P, Pajuelo E, Burgos A, Caviedes MA, et al. 2005. Isolation and characterisation of symbiotically effective *Rhizobium* resistant to arsenic and heavy metals after the toxic spill at the Aznalcóllar pyrite mine. *Soil Biology and Biochemistry* 37:1131–40
52. Ike A, Sriprang R, Ono H, Murooka Y, Yamashita M. 2007. Bioremediation of cadmium contaminated soil using symbiosis between leguminous plant and recombinant rhizobia with the *MTL4* and the *PCS* genes. *Chemosphere* 66:1670–76
53. Sá-Pereira P, Rodrigues M, Videira e Castro I, Simões F. 2007. Identification of an arsenic resistance mechanism in rhizobial strains. *World Journal of Microbiology & Biotechnology* 23:1351–56
54. Dary M, Chamber-Pérez MA, Palomares AJ, Pajuelo E. 2010. "In situ" phytostabilisation of heavy metal polluted soils using *Lupinus luteus* inoculated with metal resistant plant-growth promoting rhizobacteria. *Journal of Hazardous Materials* 177:323–30
55. Pajuelo E, Rodríguez-Llorente I, Lafuente A, Caviedes M. 2011. Legume-*Rhizobium* symbioses as a tool for bioremediation of heavy metal polluted soils. In *Biomangement of Metal-Contaminated Soils. Environmental Pollution*, eds. Khan M, Zaidi A, Goel R, Musarrat J. Volume 20. Dordrecht: Springer. pp. 95–123. [https://doi.org/10.1007/978-94-007-1914-9\\_4](https://doi.org/10.1007/978-94-007-1914-9_4).
56. Zhuang P, Yang Q, Wang H, Shu W. 2007. Phytoextraction of heavy metals by eight plant species in the field. *Water, Air, and Soil Pollution* 184:235–42
57. Greipsson S. 2011. Phytoremediation. *Nature Education Knowledge* 3:7
58. Ali H, Khan E, Sajad MA. 2013. Phytoremediation of heavy metals—concepts and applications. *Chemosphere* 91:869–81
59. Castro IV, Ferreira EM, McGrath SP. 1997. Effectiveness and genetic diversity of *Rhizobium leguminosarum* bv. *trifolii* isolates in Portuguese soils polluted by industrial effluents. *Soil Biology and Biochemistry* 29:1209–13
60. Castro IV, Ferreira EM, McGrath SP. 2003. Survival and plasmid stability of rhizobia introduced into a contaminated soil. *Soil Biology and Biochemistry* 35:49–54
61. Delorme TA, Gagliardi JV, Angle JS, van Berkum P, Chaney RL. 2003. Phenotypic and genetic diversity of rhizobia isolated from nodules of clover grown in a zinc and cadmium contaminated soil. *Soil Science Society of America Journal* 67:1746–54
62. de M Rangel W, de Oliveira Longatti SM, Ferreira PAA, Bonaldi DS, Guimarães AA, et al. 2017. Leguminosae native nodulating bacteria from a gold mine As-contaminated soil: multi-resistance to trace elements, and possible role in plant growth and mineral nutrition. *International Journal of Phytoremediation* 19:925–36
63. Estrada-de Los Santos P, Palmer M, Chávez-Ramírez B, Beukes C, Steenkamp ET, et al. 2018. Whole genome analyses suggests that *Burkholderia* sensu lato contains two additional novel genera (*Mycetohabitans* gen. nov., and *Trinickia* gen. nov.): implications for the evolution of diazotrophy and nodulation in the *Burkholderiaceae*. *Genes* 9:389
64. Deepika KV, Raghuram M, Kariali E, Bramhachari PV. 2016. Biological responses of symbiotic *Rhizobium radiobacter* strain VBCK1062 to the arsenic contaminated rhizosphere soils of mung bean. *Ecotoxicology and Environmental Safety* 134P1:1–10
65. Broos K, Uytendaele M, Mertens J, Smolders E. 2004. A survey of symbiotic nitrogen fixation by white clover grown on metal contaminated soils. *Soil Biology and Biochemistry* 36:633–40
66. Reichman SM. 2007. The potential use of the legume-rhizobium symbiosis for the remediation of arsenic contaminated sites. *Soil Biology and Biochemistry* 39:2587–93
67. Pajuelo E, Rodríguez-Llorente ID, Dary M, Palomares AJ. 2008. Toxic effects of arsenic on *Sinorhizobium-Medicago sativa* symbiotic interaction. *Environmental Pollution* 154:203–11
68. Finnegan PM, Chen W. 2012. Arsenic toxicity: the effects on plant metabolism. *Frontiers in Physiology* 3:182
69. Bustingorri C, Noriega G, Lavado RS, Balestrasse K. 2017. Protective effect exerted by soil phosphorus on soybean subjected to arsenic and fluoride. *Redox Report* 22:353–60



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